



Conservation Agriculture as Climate Change Mitigation

Mekin Mohammed Yimam

Ethiopian Institute of Agricultural Research (EIAR), Fogera National Rice Research and Training Center, P.O. Box 1937, Ethiopia

Corresponding author: mekinmmd@gmail.com

Abstract

Ethiopia's economy and ecological system are fragile and vulnerable to climate change. Conservation agriculture (CA), comprising minimum soil disturbance, retention of crop residues and crop diversification, is widely promoted for reducing soil degradation and improving agricultural sustainability. This review paper encompasses the role of conservation agriculture for climate change mitigation. Soil Organic Carbon (SOC) increases are reversible and the climate change benefit relies on the new management practice continuing indefinitely. To achieve long term climate change mitigation, the additional SOC must be in recalcitrant forms. As several studies indicate that the additional C accumulated under CA practices such as no-till is concentrated in particulate organic fractions or other forms considered 'labile' with only marginal increases in more recalcitrant form. The important point is that simply because SOC benefits are demonstrated under experimental conditions, the large scale adoption of conservation agriculture (CA) in regions dominated by resource poor smallholder farmers may be extremely slow: it is therefore unwise to rely on adoption of CA as a major strategy for climate change mitigation. One kg additional N₂O emitted per ha will offset 0.13 MG C per ha sequestered. In many cases CA practice will be functional only a small degree of climate change mitigation through soil carbon sequestration.

Keywords: Conservation Agriculture, Climate Change, Climate Change mitigation, N fertilizer

1. Introduction

1.1. Background and justification

Conservation agriculture (CA), comprising minimum soil disturbance, retention of crop residues and crop diversification, is widely promoted for reducing soil degradation and improving agricultural sustainability (Powlson et al., 2015). A group of crop management practices termed "conservation agriculture" (CA) are

widely promoted to increase crop yields, reduce soil degradation and develop systems that are more resilient to weather-induced stresses including those caused by climate variability and change (FAO, 2001; Kassam et al., 2009; Thierfelder and Wall, 2010; Jat et al., 2012). Although CA shows great promise in many agro-ecological environments, there is continuing vigorous debate about its practical feasibility under certain farmer circumstances, especially for smallholders in mixed crop/livestock systems in

tropical regions, where there is competition for crop residues between their use as animal feed as opposed to soil retention (Giller et al., 2009, 2011; Valbuena et al., 2012).

The impact of climate change on agricultural productivity is severe in Sub Saharan Africa (Mohammed *et al.*, 2020). It is due to low adoption of key production technologies that enhance adaptation to climatic change and increase productivity. Future work need to consider also studying the effects of different climate change adaptation strategies (Mohammed *et al.*, 2020). It is thus against this scenario and statistics that, rural farmers have to consequently adopt farming practices that conserve fragile soils and improve its fertility for improving crop production in marginal rainfall regions.

Ethiopia's economy and ecological system are fragile and vulnerable to climate change (César and Ekbom, 2013). Food security is highly sensitive to climate risks and rainfall is one of the main climatic determinants of food production in Ethiopia. Rainfall is highly erratic and unreliable (Stroosnijder and Van Rheenen, 2001; Mesfin et al., 2009) in respect to mainly the delay in the onset and early cessation this intermittent long dry spell throughout the growing season has a tremendous influence crop production (Rockstrom, 2000; Abdelkadir and Richard, 2005) and it is the main risk contributing to food insecurity and overall vulnerability of households. The vulnerability to climate-related hazards and food insecurity is closely linked to land degradation, in which about 85% of the land surface in Ethiopia is considered susceptible to moderate or severe soil degradation and erosion. Moreover, the main reasons for dryland cropland degradation in Ethiopia include complete removal of crop residues at harvest, aftermath overgrazing of livestock, frequent tillage, drought and inefficient use of technologies and practices (Manado, 1997; Taddese, 2001). Farmers in the study areas plough their land 3-5 times for Sorghum and maize crops per season using traditional equipment 'Maresaha'. Repeated tillage accelerates SOM decomposition (Doran and Smith, 1987) and water runoff and soil erosion

(Derpsch et al., 1991), and other manifestations of physical, chemical and biological soil degradation (Benites, 2008; Kerte' szet al., 2008), and it has been reported to be the main cause of land degradation in Ethiopia (Tefera, 2002). This review paper encompasses the role of conservation agriculture for climate change mitigation. So, the objective of this review paper was: (1) to asses soil carbon sequestration for climate change mitigation; (2) To overview soil organic carbon condition under conservation agriculture in the context of climate change mitigation; (3) To assess the adoption by farmers of practices leading to SOC increases and implications for climate change mitigation; (4) To review the Nitrogen management as a climate change mitigation.

2. Conservation agriculture as Climate change mitigation

2.1. Soil carbon sequestration for Climate change mitigation

As indicated by different scholars, examples; Batjes, (2014); Scharlemann *et al.*, (2015), the concept of sequestering organic carbon in soil to mitigate climate change arises because past land clearances caused a major decline in the global SOC stock and released CO₂ to the atmosphere. so, in principle, there is potential to reverse this trend (Smith *et al.*, 2008). SOC stocks are significant at the global scale, recently estimated at some 1500 Pg C to a depth of 1 m (Batjes, 2014; Scharlemann *et al.*, 2015), about twice the amount in atmospheric CO₂ and three times that in global vegetation. However, there are well known general limitations to the approach (Freibauer *et al.*, 2004; Janzen, 2006; Powlson *et al.*, 2011a; Stockmann *et al.*, 2013; Sommer and Bossio, 2014).

According to Dimassiet *al.* (2013), the capacity for SOC accumulation is finite. Carbon content moves towards a new equilibrium value then increases no further. SOC increases are reversible and the climate change benefit relies on the new management practice continuing indefinitely. To

deliver long-term climate change mitigation the additional SOC must be in recalcitrant forms. Several studies indicate that the additional C accumulated under CA practices such as no-till is concentrated in particulate organic fractions or other pools considered 'labile', with only marginal increases in more recalcitrant pools (Wander *et al.*, 1998; Machado *et al.*, 2003; Chivenge *et al.*, 2007; Bhattacharyya *et al.*, 2011, 2012; Tivet *et al.*, 2013; O'Rourke *et al.*, 2015). The largest amounts of SOC sequestration are obtained by taking agricultural land out of production and returning it to native grassland or forest but this change in land use conflicts with meeting food security goals (Smith *et al.*, 2008).

In addition, reliably monitoring relatively small changes in Soil Organic Carbon (SOC) in agricultural soils following a management change is challenging due to a combination of soil variability and slow rates of change (Smith, 2004; Batjes and van Wesemael, 2015). This is exacerbated in many tropical regions due to a lack of monitoring networks or long-term experiments to complement site-based measurements (Batjes and van Wesemael, 2015). On the positive side, a soil C sequestration strategy can, in principle, start immediately, is not dependent on development of new technologies and will normally improve soil quality and contribute to a range of ecosystem services (Verhulst *et al.*, 2010; Powlson *et al.*, 2011b, 2014; Palm *et al.*, 2013; Lal, 2015). Even small increases in labile SOC fractions can have positive effects on a range of soil physical properties, and thus potentially improve the resilience of soils to stress and contribute to climate change adaptation (Watts *et al.*, 2006; Powlson *et al.*, 2011b; Thierfelder and Wall, 2012; Thierfelder *et al.*, 2013a; Chakraborty *et al.*, 2014).

As indicated by different scholars, examples; Batjes, (2014); Scharlemann *et al.*, (2015), the concept of sequestering organic carbon in soil to mitigate climate change arises because past land clearances caused a major decline in the global SOC stock and released CO₂ to the atmosphere. so, in principle, there is potential to reverse this trend (Smith *et al.*, 2008). SOC stocks are significant at the global scale, recently estimated

at some 1500 Pg C to a depth of 1 m (Batjes, 2014; Scharlemann *et al.*, 2015), about twice the amount in atmospheric CO₂ and three times that in global vegetation. However, there are well known general limitations to the approach (Freibauer *et al.*, 2004; Janzen, 2006; Powlson *et al.*, 2011a; Stockmann *et al.*, 2013; Sommer and Bossio, 2014).

According to Dimassi *et al.* (2013), the capacity for SOC accumulation is finite. Carbon content moves towards a new equilibrium value then increases no further. SOC increases are reversible and the climate change benefit relies on the new management practice continuing indefinitely. To deliver long-term climate change mitigation the additional SOC must be in recalcitrant forms. Several studies indicate that the additional C accumulated under CA practices such as no-till is concentrated in particulate organic fractions or other pools considered 'labile', with only marginal increases in more recalcitrant pools (Wander *et al.*, 1998; Machado *et al.*, 2003; Chivenge *et al.*, 2007; Bhattacharyya *et al.*, 2011, 2012; Tivet *et al.*, 2013; O'Rourke *et al.*, 2015). The largest amounts of SOC sequestration are obtained by taking agricultural land out of production and returning it to native grassland or forest but this change in land use conflicts with meeting food security goals (Smith *et al.*, 2008).

In addition, reliably monitoring relatively small changes in Soil Organic Carbon (SOC) in agricultural soils following a management change is challenging due to a combination of soil variability and slow rates of change (Smith, 2004; Batjes and van Wesemael, 2015). This is exacerbated in many tropical regions due to a lack of monitoring networks or long-term experiments to complement site-based measurements (Batjes and van Wesemael, 2015). On the positive side, a soil C sequestration strategy can, in principle, start immediately, is not dependent on development of new technologies and will normally improve soil quality and contribute to a range of ecosystem services (Verhulst *et al.*, 2010; Powlson *et al.*, 2011b, 2014; Palm *et al.*, 2013; Lal, 2015). Even small increases in labile SOC fractions can have positive effects on a range of soil physical properties, and thus

potentially improve the resilience of soils to stress and contribute to climate change adaptation (Watts *et al.*, 2006; Powlson *et al.*, 2011b; Thierfelder and Wall, 2012; Thierfelder *et al.*, 2013a; Chakraborty *et al.*, 2014).

2.2. Soil organic carbon increase under conservation agriculture in the context of climate change mitigation.

Many conservation agriculture (CA) practices including zero tillage increase the concentration of soil organic Carbon (SOC) in the surface layers of soil, which is beneficial for soil quality, but it does not necessarily follow that SOC stock is increased which is necessary for climate change mitigation. According to Mohammed (2022) the mean SOC stocks were higher among the vegetation communities in Northern than in the Southern aspects of Bale Mountain. CA contributes for vegetation community increment. Furthermore, it is incorrect to assume that every management practice leading to an increase in SOC stock mitigates climate change: interpretation has to take account of the processes leading to the SOC increase and whether or not there is a net transfer of C from atmosphere to soil compared to conventional practice (Powlson *et al.*, 2011a; Stockmann *et al.*, 2013; Sommer and Bossio, 2014). In the context of CA, certain types of crop diversification undoubtedly deliver an increased transfer of C from atmosphere to soil through a greater input of photosynthate. This is true for legume inter-crops grown between rows of maize, as promoted in regions of SSA where rainfall is sufficient to support growth of the extra crop without compromising yield of the main crop (Sakala *et al.*, 2000; Myaka *et al.*, 2006; Thierfelder *et al.*, 2013c) or where an additional crop is grown in the period between other crops where the soil would otherwise be fallow (double or relay cropping; Ghosh *et al.*, 2012; Mupangwa and Thierfelder, 2014). It is also possible that replacing one of the crops in an existing system with another could increase C inputs to soil: this depends on the root biomass of the replacement crop and the proportion of its above-ground biomass returned to soil compared to the conventional system. It also depends on the

decomposability of the input from the replacement crop as influenced by its chemical composition (Vanlauwe *et al.*, 2005). In addition to increasing soil C stock and concentration, diversification can deliver benefits to the farmer through the economic value of the additional crop, though there are often significant barriers to overcome regarding availability of seed, additional labour requirements and the establishment of local markets for new crops (Thierfelder *et al.*, 2012, 2013c). In principle, zero tillage can lead to genuine C sequestration and climate change mitigation if it slows the rate of decomposition of existing SOC or promotes stabilization of incoming organic C (Powlson *et al.*, 2011a; Stockmann *et al.*, 2013), but the impact on SOC is often a matter of depth redistribution rather than net accumulation, at least in the short to medium term (Baker *et al.*, 2007; Angers and Eriksen-Hamel, 2008; Govaerts *et al.*, 2009; Luo *et al.*, 2010; Virto *et al.*, 2012; Powlson *et al.*, 2014).

The studies in temperate region showed that the impact of cereal straw incorporation on SOC content to be small and often non-significant, even when continued for up to 25 years (Powlson *et al.*, 2011b). The rate of residue decomposition, and hence SOC accumulation, is more sensitive to environmental conditions (temperature, moisture) for surface-applied residues, as in CA, than for those that are incorporated (Helgason *et al.*, 2014). Smaller SOC increases are expected in tropical regions compared to temperate due to more rapid decomposition under higher mean temperatures. Whether or not SOC increases produced by residue retention constitute climate change mitigation depends on the alternate fate of the residues under conventional practice. If the alternative is burning in the field, as is often the case in rice wheat systems in the western IGP, then any retention of residue-derived C in soil in a CA treatment, however small, is C that would otherwise have been emitted to the atmosphere during burning: in this situation the SOC increase represents genuine climate change mitigation. Residue-derived C not respired during digestion by the animal will be converted into manure and in most situations, certainly in SSA, manure is

applied to soil by farmers because it is a valuable source of crop nutrients and of organic matter for improving soil properties. There will be some gaseous loss of manure C to the atmosphere as CO₂ or methane, the amounts depending on numerous factors especially the period of manure storage and whether or not it is composted. But at least part of the original residue-derived C would have been returned to soil in manure under conventional practice, though often to a different field. Consequently, it is incorrect to treat the whole of the SOC increase from residue retention under CA as climate change mitigation. Because of the diverse range of situations in the two regions it is not possible to make a realistic estimate of the proportion of residue-derived C that would be returned under conventional practice, and hence the appropriate amount to discount SOC increases from direct residue return under CA. A significant quantity of animal manure (which includes residue-derived C) is dried and then burned as fuel for cooking or domestic heating, leading to complete loss of its C to the atmosphere. In these situations, it is reasonable to regard SOC increases from direct return of crop residues as retention of C that would otherwise have been emitted to the atmosphere and thus genuine climate change mitigation. Provision of alternative domestic energy sources in rural areas would be an influential policy to promote residue retention in the field with benefits for both climate change mitigation and soil quality (Helgason et al., 2014).

2.3. Adoption by farmers of practices leading to SOC increases and implications for climate change mitigation

Even where genuine increases in SOC from CA can be demonstrated in experiments, and where they can correctly be interpreted as delivering climate change mitigation through soil C sequestration, there are often significant technical, infrastructural, social or policy barriers to the adoption of the new practices by smallholder farmers in regions such as SSA. This topic has been discussed in detail elsewhere (Gowing and Palmer, 2008; Govaertset al., 2009; Giller et al., 2011; Thierfelder et al., 2012; Andersson and

D'souza, 2013; Thierfelder et al., 2014; Corbeels et al., 2014; Mason et al., 2015) and there are some examples of these barriers being overcome (Thierfelder et al., 2014; Titttonellet al., 2012). The important point is that simply because SOC benefits are demonstrated under experimental conditions, the large-scale adoption of CA in regions dominated by resource-poor smallholder farmers may be extremely slow: it is therefore unwise to rely on adoption of CA as a major strategy to mitigate climate change.

2.4. Nitrogen management as climate change mitigation strategy as compared to conservation agriculture.

To place the degree of climate change mitigation possible from CA into a wider context it is appropriate to compare with the potential from other agricultural approaches. Management of nitrogen (N) fertilizer is an obvious comparison because of its large greenhouse gas footprint (Brentrup and Pallière, 2008; Zhang et al., 2013; Cheng et al., 2015). There are numerous examples of relatively simple improvements in N fertilizer management either leading to decreased total applications whilst maintaining, or even increasing, crop yields or of changes in practice leading to increased N use efficiency such that a given rate of N produces increased yield (Khurana et al., 2008; Sapkota et al., 2014). These improvements in N management inevitably lead to decreased emissions of N₂O or, even if total emission is not decreased, to a decrease in yield-scaled emissions (Bhatia et al., 2012; Chauhan et al., 2012). i.e., emission expressed per ton of grain (Linguist et al., 2012). As indicated by Davison, (2009), an emission saving equivalent to 0.16 Mg C ha⁻¹ yr⁻¹ if calculated using an emission factor of 2.5% for N fertilizer (i.e., 2.5% of applied N fertilizer released to the atmosphere as N₂O-N through a combination of direct plus indirect emissions (Davidson, 2009). This is of the same order as GHG reductions possible from soil C sequestration. According to Smith et al. (2012), each 1 kg of N fertilizer is responsible for an emission of 11.7 kg CO₂e, equivalent to a change of 3.2 kg SOC. The benefit from N management could be even greater as there is

some evidence to indicate that emission factors could be as large as 4% of applied N (Smith *et al.*, 2012). Even using the current IPCC default emission factor of 1.25% the GHG saving from reduced N fertilizer use is substantial and sustainable. Even relatively small changes in the timing of N applications, but with the same total dose, can deliver a significant decrease in N₂O emission. The measured annual N₂O emission decreased by 16%, equivalent to 0.04 Mg C ha⁻¹ (Bhatia *et al.*, 2012). A key advantage of decreasing GHG emissions through N management is that unlike soil C sequestration the benefits are irreversible and can continue indefinitely. In addition, CA practices themselves can sometimes increase the efficiency of use of N fertilizer (Jat *et al.*, 2012; Sapkota *et al.*, 2014), thus providing a synergy between CA and N fertilizer management that contributes to decreased GHG emissions. There are however numerous barriers to increased inputs of fertilizers due to a range of cost and infrastructure issues (Montpellier Panel, 2014) so any conservation of nutrients through improved management such as retention of crops residues, or inputs of N through inclusion of legumes as an aspect of crop diversification, is highly beneficial for food security whether or not it contributes to climate change mitigation. Increased N inputs from any source will almost inevitably cause an increase in total N₂O emission from agriculture in the SSA. This would appear to be a necessary cost of improving food security however, with the associated increased yields, yield-scaled emissions would be expected to decrease (Linguist *et al.*, 2012).

2.5. Effect of reduced tillage on nitrous oxide emissions

In some circumstances zero tillage may lead to increased emission of N₂O, but this is not universally observed (Van Kessel *et al.*, 2013). There is currently a paucity of evidence on the impact of zero tillage or other aspects of CA on N₂O emissions in the IGP or SSA regions. However, the issue is crucial for determining the potential for CA to mitigate climate change as even a small increase in N₂O emission will offset

a gain in SOC (Grandy *et al.*, 2006). One kg additional N₂O emitted per ha will offset 0.13 Mg C ha⁻¹ sequestered. In a situation with high rates of N fertilizer, a combination of no-till and straw retention led to a decrease in N₂O emission but equal or increased crop yields compared to conventional tillage with straw removed (Huang *et al.*, 2015). By contrast, a study in Madagascar with intercropped maize-soybean on a clay soil (Chapuis-Lardy *et al.*, 2009) showed no difference in N₂O emission between a direct seeded mulch based system and traditional hand ploughing with previous crop residues removed.

3. Conclusion and recommendation

As different literatures showed, globally CA contribute significant for food security, and shows a general trend for all of the practices constituting CA to cause some increase in SOC stock, implying some degree of climate change mitigation through soil C sequestration. The wide variability between sites and agricultural systems indicates that it is incorrect to assume that results from a specific location are necessarily transferable. Most of the published values are almost certainly inflated by 20–30% (Palm *et al.*, 2013) due to the inappropriate soil sampling method used: equal soil depth rather than equal soil mass. This is an important yet frequently overlooked point. Where a significant increase in SOC stock is measured it is essential to assess the specific management practice responsible in order to determine whether it represents a net additional transfer of C from atmosphere to land, and hence genuine climate change mitigation, rather than a spatial redistribution of organic C in soil. SOC increases caused by crop diversification will almost certainly constitute genuine mitigation; this third principle of CA, which is often overlooked, warrants further assessment, especially in view of the potential multiple benefits it can deliver. However, it must be recognized that adoption of this change, and other CA practices, face considerable practical barriers for resource-poor smallholder farmers in the regions we considered. Thus, it would be unwise to assume that CA can be a key strategy for

climate change mitigation. Therefore, we can conclude that in many cases CA practices will deliver only a small degree of climate change mitigation through soil carbon sequestration. CA practices can reasonably be regarded as contributing to climate change adaptation and to sustainable intensification, whether or not they consistently deliver increased crop yields in every season. We suggest that CA should be promoted on these grounds, in addition to any more direct livelihood benefits to farmers that will vary between regions and economic situations. Any contribution to climate change

mitigation should be regarded as a welcome additional benefit, not as a key policy driver for promoting the practices. From a policy perspective there is a risk that a narrow focus on soil C sequestration can lead to an exaggerated view of the opportunities for climate change mitigation through CA or related practices with too little attention given to other approaches that may have greater potential and be more easily achieved in practice. One such example is the management of N fertilizer, whether through its direct management or through ensuring an adequate supply of other nutrients.

4. References

- Abdelkdair, A. and Schultz, R.C., 2005. Water harvesting in a 'runoff-catchment' agroforestry system in the dry lands of Ethiopia. *Agroforestry systems*, 63(3), pp.291-298.
- Andersson, J.A., D'souza, S., 2013. From adoption claims to understanding farmers and contexts: a literature review of Conservation Agriculture (CA) adoption among smallholder farmers in southern Africa. *Agric. Ecosyst. Environ.* 187, 116–132.
- Angers, D.A., Eriksen-Hamel, N.S., 2008. Full-inversion tillage and organic carbon distribution in soil profiles: a meta-analysis. *Soil Sci. Soc. Am. J.* 72, 1370–1374.
- Baker, J.M., Ochsner, T.E., Venterea, R.T., Griffis, T.J., 2007. Tillage and soil carbon sequestration—what do we really know? *Agric. Ecosyst. Environ.* 118, 1–5.
- Batjes, N.H., 2014. Total carbon and nitrogen in the soils of the world. *Eur. J. Soil Sci.* 65, 4–21.
- Batjes, N.H., 2014. Total carbon and nitrogen in the soils of the world. *Eur. J. Soil Sci.* 65, 4–21.
- Batjes, N.H., van Wesemael, B., 2015. Measuring and monitoring soil carbon. In: Banwart, S.A., Noellemeier, E., Milne, E. (Eds.), *Soil Carbon Science, Management and Policy for Multiple Benefits*. SCOPE Series, vol. 71. CABI, Wallingford, UK, pp. 188–201.
- Benites, J., 2008. Effect of no-tillage on conservation of the soil and soil fertility. In: Goddard, T., Zöbisch, M.A., Gan, Y.T., Ellis W, Watson, A., Sombatpanit, S. (Eds.).
- Bhatia, A., Pathak, H., Jain, N., Singh, P.K., Tomer, R., 2012. Greenhouse gas mitigation in rice–wheat system with leaf color chart-based urea application. *Environ. Monit. Assess.* 184, 3095–3107.
- Bhattacharyya, R., Kundu, S., Srivasta, A.K., Gupta, H.S., Prakesh, V., Bhatt, J.C., 2011. Long term fertilization effects on soil organic carbon pools in a sandy soil of the Indian sub-Himalayas. *Plant Soil* 341, 109–124.
- Bhattacharyya, R., Tuti, M., D, Kundu, S., Bisht, J.K., Bhatt, J.C., 2012. Conservation tillage Impacts on soil aggregation and carbon pools in a sandy clay loam soil of the Indian Himalayas. *Soil Sci. Soc Am. J.* 76, 617–627.
- Brentrup, F., Pallière, C., 2008. GHG emissions and energy efficiency in European nitrogen fertiliser production and use. *Proceedings 639, International Fertiliser Society* 1–25.

- César, E. and Ekbom, A., 2013. Ethiopia environmental and climate change policy brief. *Sida's Helpdesk for Environment and Climate Change*.
- Chakraborty, D., Watts, C., W, Powlson, D.S., Macdonald, A.J., Ashton, R.W., White, R. P., Whalley, W.R., 2014. Triaxial testing to determine the effect of soil type and organic carbon content on soil consolidation and shear deformation characteristics. *Soil Sci. Soc. Am. J.* 78, 1192–1200.
- Chapuis-Lardy, L., Metay, A., Martinet, M., et al., 2009. Nitrous oxide fluxes from Malagasy agricultural soils. *Geoderma* 148, 421–427.
- Chauhan, B.S., Mahajan, G., Sardana, V., Timsina, J., Jat, M.L., 2012. Productivity and sustainability of the rice–wheat cropping system in the Indo-Gangetic Plains of the Indian subcontinent: problems opportunities, and strategies. *Adv. Agron.* 117, 315–369.
- Cheng, K., Yan, M., Nayak, D., Pan, D.X., Smith, P., Zheng, J.F., Zheng, J.W., 2015. Carbon footprint of crop production in China: an analysis of national statistics data. *J. Agric. Sci.* 153, 422–431.
- Chivenge, P.P., Murwira, H.K., Giller, K.E., Mapfumo, P., Six, J., 2007. Long-term impact of reduced tillage and residue management on soil carbon stabilization: implications for conservation agriculture on contrasting soil. *Soil Till. Res.* 94, 328–337.
- Corbeels, M., Sakyi, R.K., Kühne, R.F., Whitbread, A. (2014) Meta-analysis of crop responses to conservation agriculture in sub-Saharan Africa. CCAFS Report No. 12. Copenhagen: CGIAR Research Program on Climate Change, Agriculture and Food Security (CAAFS).
- Davidson, E.A., 2009. The contribution of manure and fertilizer nitrogen to atmospheric nitrous oxide since 1860. *Nat. Geosci.* 2, 659–662.
- Derpsch, R., Roth, C., Sidiras, N., Köpcke, U., 1991. Controle de erosão no Paraná, Brasil: Sistemas de cobertura do solo, plantio direto e preparo conservacionista do solo. Deutsche Gesellschaft für Technische Zusammenarbeit (GTZ) GmbH, Eschborn, Germany.
- Dimassi, B., Cohan, J.-P., Labreuche, J., Mary, B., 2013. Changes in soil carbon and nitrogen following tillage conversion in a long-term experiment in Northern France. *Agric. Ecosyst. Environ.* 169, 12–20.
- Doran, J.W., Smith, M.S., 1987. Organic matter management and utilization of soil and fertilizer nutrients. *Soil Fertility and Organic Matter as Critical Components of Production Systems*, SSSA Special Publication No. 19, Soil Science Society Inc./American Society of Agronomy, Inc. Publ., Madison, WI, pp.53–72.
- FAO, 2001. Conservation Agriculture Case Studies in Latin America and Africa. Introduction. *FAO Soils Bulletin* No. 78. FAO, Rome.
- Freibauer, A., Rounsevell, M.D.A., Smith, P., Verhagen, J., 2004. Carbon sequestration in the agricultural soils of Europe. *Geoderma* 122, 1–23.
- Ghosh, P.K., Venkatesh, M.S., Hazra, K.K., Narendra, K., 2012. Long-term effect of pulses and nutrient management on soil organic carbon dynamics and sustainability on an Inceptisol of Indo-Gangetic Plains of India. *Exp. Agric.* 48, 473–487.
- Giller, K.E., Corbeels, M., Nyamangara, J., Triomphe, B., Affholder, F., Scopel, E., Tittonell, P., 2011. A research agenda to explore the role of conservation agriculture in African smallholder farming systems. *Field Crop Res.* 124, 468–472.
- Giller, K.E., Witter, E., Corbeels, M., Tittonell, P., 2009. Conservation agriculture and smallholder farming in Africa: the heretics view. *Field Crop Res.* 114, 23–34.

- Govaerts, B., Verhulst, N., Castellanos-Navarrete, A., Sayre, K.D., Dixon, J., Dendooven, L., 2009. Conservation agriculture and soil carbon sequestration: between myth and farmer reality. *Crit. Rev. Plant Sci.* 28, 97–122.
- Gowing, J.W., Palmer, M., 2008. Sustainable agricultural development in sub-Saharan Africa: the case for a paradigm shift in land husbandry. *Soil Use Manage.* 24, 92–99.
- Grandy, A.S., Loecke, T.D., Parr, S., Robertson, G.P., 2006. Long-term trends in nitrous oxide emissions, soil nitrogen, and crop yields of till and no-till cropping systems. *J. Environ. Sci.* 35, 1487–1495.
- Helgason, B.L., Gegerich, E.G., Janzen, H.H., Ellert, B.H., Lorenz, N., Dick, R.P., 2014. Long-term microbial retention of residue is site-specific and depends on residue placement. *Soil Biol. Biochem.* 68, 231–240.
- Huang, M., Liang, T., Wang, L., 2015. Nitrous oxide emissions in a winter wheat–summer maize double cropping system under different tillage and fertilizer management. *Soil Use Manage.* 31, 98–105.
- Janzen, H.H., 2006. The soil carbon dilemma: shall we hoard it or use it? *Soil Biol. Biochem.* 38, 419–424.
- Jat, R.A., Wani, S.P., Sahrawat, K.L., 2012. Conservation Agriculture in the semi-arid tropics: prospects and problems. *Adv. Agron.* 117, 191–273.
- Kassam, A., Friedrich, T., Shaxson, F., Pretty, J., 2009. The spread of conservation agriculture: justification, sustainability and uptake. *Int. J. Agric. Sustain.* 7, 292–320.
- Khurana, H.S., Phillips, S.B., Singh, B., et al., 2008. Agronomic and economic evaluation of site specific nutrient management for irrigated wheat in northwest India. *Nutr. Cycl. Agroecosyst.* 82, 15–31.
- Lal, R., 2015. Sequestering carbon and increasing productivity by conservation agriculture. *J. Soil Water Conserv.* 70, 55A–62A.
- Linquist, B., van Groenigen, K.J., Adviento-Borbe, M.A., Pittelkow, C., van Kessel, C., 2012. An agronomic assessment of greenhouse gas emissions from major cereal crops. *Glob. Change Biol.* 18, 194–209.
- Luo, Z., Wang, E., Sun, O.J., 2010. Can no-tillage stimulate carbon sequestration in agricultural soils? A meta-analysis of paired experiments. *Agr. Ecosyst. Environ.* 139, 224–231.
- Machado, P.L.O.A., Sohi, S.P., Gaunt, J.L., 2003. Effect of no-tillage on turnover of organic matter in a Rhodic Ferralsol. *Soil Use Manage.* 19, 250–256.
- Mando, A., 1997. The effect of mulch on the water balance of Sahelian crusted-soils. *Soil Technology*, 11, pp.121-138.
- Mason, S.C., Ouattara, K., Taonda, S.J., Pale', S., Sohero, A., Kabore', D., 2015. Soil and cropping system research in semi-arid West Africa as related to the potential for conservation agriculture. *Int. J. Agric. Sustain.* 13, 120–134.
- Mesfin, T., Tesfahunegn, G.B., Wortmann, C.S., Nikus, O. and Mamo, M., 2009. Tied-ridging and fertilizer use for sorghum production in semi-arid Ethiopia. *Nutrient cycling in agroecosystems*, 85(1), pp.87-94.
- Mohammed, M., Biazn, B. and Belete, M.D., 2020. Hydrological impacts of climate change in TikurWuha watershed, Ethiopian Rift Valley Basin. *J Environ Earth Sci*, 10(2), pp.28-49.
- Mohammed M. 2022. The influence of topographic aspect on biomass and soil carbon stock in case of Bale Mountain National Park, Ethiopia. *J Environ Earth Sci*, 12(2), pp.5-7.

- Montpellier Panel 2014. The Montpellier Panel, December 2014. No Ordinary Matter: conserving, restoring and enhancing Africa's soils.
- Mupangwa, W., Thierfelder, C., 2014. Intensification of conservation agriculture systems for increased livestock feed and maize production in Zimbabwe. *Int. J. Agric. Sustain.* 12, 425–439.
- Myaka, F.M., Sakala, W.D., Adu-Gyamfi, J.J., et al., 2006. Yields and accumulations of N and P in farmer-managed intercrops of maize-pigeonpea in semi-arid Africa. *Plant Soil* 285, 207–220.
- O'Rourke, S.M., Angers, D.A., Holden, N.M., McBratney, A.B., 2015. Soil organic carbon across scales. *Glob. Change Biol.* 21, 3561–3574.
- Palm, C., Blanco-Canqui, H., DeClerck, F., Gatere, L., 2013. Conservation agriculture and ecosystem services: an overview. *Agric. Ecosyst. Environ.* 187, 87–105.
- Powlson David S, Stirling Clare M., Thierfelder Christian, White Rodger P., Jat M.L. (2015). Does conservation agriculture deliver climate change mitigation through soil carbon sequestration in tropical agroecosystems? *Agriculture, Ecosystems and Environment* 220 (2016) 164–174.
- Powlson, D.S., Glendining, M.J., Coleman, K., Whitmore, A.P., 2011b. Implications for soil properties of removing cereal straw: results from long-term studies. *Agron. J.* 103, 279–287.
- Powlson, D.S., Stirling, C.M., Jat, M.L., Gerard, B.G., Palm, C.A., Sanchez, P.A., Cassman, K.G., 2014. Limited potential of no-till agriculture for climate change mitigation. *Nat. Clim. Change* 4, 678–683.
- Powlson, D.S., Whitmore, A.P., Goulding, K.W.T., 2011a. Soil carbon sequestration to mitigate climate change: a critical re-examination to identify the true and the false. *Eur. J. Soil Sci.* 62, 42–55.
- Rockstrom, J., 2000. Water resources management in smallholder farms in Eastern and Southern Africa: an overview. *Physics and Chemistry of the Earth, Part B: Hydrology, Oceans and Atmosphere*, 25(3), pp.275–283.
- Sakala, W.D., Cadisch, G., Giller, K.E., 2000. Interactions between residues of maize and pigeonpea and mineral N fertilizers during decomposition and N mineralization. *Soil Biol. Biochem.* 32, 679–688.
- Sapkota, T.B., Majumdar, K., Jat, M.K., Kumar, A., Bishnoi, D.K., McDonald, A.J., Pampolino, M., 2014. Precision nutrient management in conservation agriculture based wheat production of Northwest India: profitability, nutrient use efficiency and environmental footprint. *Field Crop. Res.* 155, 233–244.
- Scharlemann, J.P.W., Tanner, E.V.J., Hiederer, R., Kapos, V., 2015. Global soil carbon: understanding and managing the largest terrestrial carbon pool. *Carbon Manage.* 5, 81–91.
- Smith, K.A., Mosier, A.R., Crutzen, P.J., Winiwarter, W., 2012. The role of N₂O derived from crop-based biofuels, and from agriculture in general, in Earth's climate. *Philos. Trans. R. Soc. Lond. B* 367, 1169–1174.
- Smith, P., 2004. How long before a change in soil organic carbon can be detected? *Glob. Change Biol.* 10, 1878–1883.
- Smith, P., Martino, D., Cai, Z., et al., 2008. Greenhouse gas mitigation in agriculture. *Philos. Trans. R. Soc. Lond. B* 363, 789–813.
- Sommer, R., Bossio, D., 2014. Dynamics and climate change mitigation potential of soil organic carbon sequestration. *J. Environ. Manage.* 144, 83–87.

- Stockmann, U., Adams, M.A., Crawford, J.W., et al., 2013. The knowns, known unknowns and unknowns of sequestration of soil organic carbon. *Agr. Ecosyst. Environ.* 164, 80–99.
- Stroosnijder, L. and van Rheenen, T., 2001. Agro-silvo-pastoral land use in Sahelian villages.
- Taddese, G., 2001. Land degradation: a challenge to Ethiopia. *Environmental management*, 27(6), pp.815-824.
- Tefera, B., 2002. *Nature and causes of land degradation in the Oromiya Region: A review* (Vol. 36). ILRI (aka ILCA and ILRAD).
- Thierfelder, C., Cheesman, S., Rusinamhodzi, L., 2013c. Benefits and challenges of crop rotations in maize-based conservation agriculture (CA) cropping systems of southern Africa. *Int. J. Agric. Sustain.* 11, 108–124.
- Thierfelder, C., Chisui, J.L., Gama, M., et al., 2013a. Maize-based conservation agriculture systems in Malawi: long-term trends in productivity. *Field Crop. Res.* 142, 47–57.
- Thierfelder, C., Rusinamhodzi, R., Ngwira, A.R., Mupangwa, W., Nyagumbo, I., Kassie, G.T., Cairns, J.E., 2014. Conservation agriculture in Southern Africa: advances in knowledge. *Renew. Agric. Food Syst.* 30, 328.
- Thierfelder, C., Wall, P.C., 2012. Effects of conservation agriculture on soil quality and productivity in contrasting agro-ecological environments of Zimbabwe. *Soil Use Manage.* 28, 209–220.
- Tittonell, P., Scopel, E., Andrieu, E., 2012. Agroecology-based aggradation conservation agriculture (ABACO): targeting innovations to combat soil degradation and food insecurity in semi-arid Africa. *Field Crop. Res.* 132, 168–174.
- Tivet, F., Sá, J.C.M., Lal, R., Borszowskei, P.R., Briedis, C., et al., 2013. Soil organic carbon fraction loss upon continuous plow-based tillage and its restoration by diverse biomass-C inputs under no-till in sub-tropical and tropical regions of Brazil. *Geoderma* 209–210 214–225.
- Valbuena, D., Erenstein, O., Tui, S.H., et al., 2012. Conservation Agriculture in mixed crop livestock systems: scoping crop residue trade-offs in Sub-Saharan Africa and South Asia. *Field Crop. Res.* 132, 175–184.
- Van Kessel, C., Venterea, R., Six, J., et al., 2013. Climate duration, and N placement determine N₂O emissions in reduced tillage systems: a meta-analysis. *Global Change Biol.* 19, 3336.
- Vanlauwe, B., Gachengo, C., Shepherd, K., Barrios, E., Palm, C., 2005. Laboratory validation of a resource quality-based conceptual framework for organic matter management. *Soil Sci. Soc. Am. J.* 69, 1135–1145.
- Verhulst, N., Govaerts, B., Verachtert, E., et al., 2010. Conservation Agriculture, Improving Soil Quality for Sustainable Production Systems. In: Lal, R., Stewart, B. A. (Eds.), *Advances in Soil Science: Food Security and Soil Quality*. CRC Press, Boca Raton, FL, USA, pp. 137–208.
- Virto, I., Burlot, P., Chenu, C., 2012. Carbon input differences as the main factor explaining the variability in soil organic C storage in no-tilled compared to inversion tilled agrosystems. *Biogeochemistry* 108, 17–26.
- Wander, M.M., Bidart, M.G., Aref, S., 1998. Tillage impacts on depth distribution of total and particulate organic matter in three Illinois soils. *Soil Sci. Soc. Am. J.* 62, 1704–1711.

Watts, C.W., Clark, L.J., Poulton, P.R., Powlson, D.S., Whitmore, A.P., 2006. The role of clay, organic carbon and long-term management on mouldboard plough draught measured on the Broadbalk wheat experiment at Rothamsted. *Soil Use Manage.* 22, 334–341.

Zhang, W.F., Dou, Z.X., He, P., et al., 2013. New technologies reduce greenhouse gas emissions from nitrogenous fertilizer in China. *Proc. Natl. Acad. Sci. U. S. A.* 110, 8375–8380.

Access this Article in Online	
	Website: www.ijarbs.com
	Subject: Climatology
Quick Response Code	
DOI: 10.22192/ijarbs.2022.09.05.016	

How to cite this article:

Mekin Mohammed Yimam. (2022). Conservation Agriculture as Climate Change Mitigation. *Int. J. Adv. Res. Biol. Sci.* 9(5): 144-155.

DOI: <http://dx.doi.org/10.22192/ijarbs.2022.09.05.016>